

CHANGES IN THE FLORISTIC COMPOSITION OF A *TERRA FIRME* RAIN FOREST IN BRAZILIAN AMAZONIA OVER AN EIGHT-YEAR PERIOD IN RESPONSE TO LOGGING.¹

João Olegário Pereira de CARVALHO²

ABSTRACT - Changes in the floristic composition over an eight-year period in a logged area at the Tapajós National Forest in Brazilian Amazonia are discussed. Two treatments of different intensities of logging were compared with an undisturbed (control) forest. Data were collected from permanent sample-plots. The effects of logging on floristic composition were stronger in the more heavily logged treatment. The number of species decreased immediately after logging, but started to increase before the fifth year after logging and was higher at the end of the study period than before logging. The more heavily logged plots responded more to disturbances, as judged by the increase in the number of species during the period after logging. This forest appears to recover its initial floristic composition after disturbance without intervention.

Key-words: Amazonian forest, floristic composition, forest dynamics, logging, Tapajós Forest.

Mudanças na Composição Florística de Uma Floresta Chuvosa de Terra Firme na Amazônia Brasileira durante Um Período de Oito Anos em Resposta à Exploração Mecanizada.

RESUMO - São discutidas as mudanças na composição florística de uma área na Floresta Nacional do Tapajós, na Amazônia brasileira, no período de oito anos. Foram estudadas duas intensidades de exploração e comparadas com uma floresta não-explorada. Os dados foram coletados em parcelas permanentes. Maiores mudanças ocorreram no tratamento com a intensidade de exploração mais pesada. O número de espécies decresceu imediatamente após a exploração, porém começou a crescer cinco anos depois e foi maior no final do período estudado do que antes da exploração. As parcelas onde a exploração foi mais pesada responderam melhor aos distúrbios, a julgar pelo aumento no número de espécies durante o período pós-exploratório. A floresta parece ser capaz de recuperar natural e rapidamente sua composição florística inicial.

Palavras-chave: Floresta amazônica, composição florística, dinâmica da floresta, exploração florestal, Floresta do Tapajós.

Introduction

Floristic composition is one of the first aspects to be analyzed in all forest areas, which are the subject of ecological research, silvicultural management, or any other use. It is essential to study the floristic composition in order to evaluate the forest structure, species diversity, spatial distributions of

species, species dynamics and the growth rates of species. The work reported here is such a basis for the ecological and silvicultural research being conducted by Embrapa (Brazilian Agricultural Research Corporation) in a 180 ha area at the Tapajós National Forest in Brazilian Amazonia.

Some studies already carried out in the area are described in Carvalho

¹This paper is part of the "Forest Structure Project" (08.2000.024), Embrapa Amazônia Oriental
²Forest Researcher, Ph.D., Embrapa Amazônia Oriental, Caixa Postal 48, Belém, PA, Brazil, CEP 66.017-930. E-mail: olegario@cpatu.embrapa.br

(1992). These include: a sampling of trees ≥ 5 cm dbh to determine the basal area ha^{-1} , an inventory of trees 15–45 cm dbh to provide information about the potential stocking for the future, and a pre-logging inventory of trees ≥ 45 cm dbh (Silva *et al.*, 1985); a study on sociability of commercial species by Lopes *et al.* (1986); a commercial logging trial and a study on the reduction of basal area at the two intensities of logging according to the treatments of the present study (Carvalho *et al.*, 1986); and a preliminary study of costs and production of logging operations (Costa & Lopes, 1982). Further information about research undertaken in the study area are found in Carvalho (1984; 1992), Carvalho *et al.* (1984, 1993) and Lopes (1993). The present paper is part of a series of papers dealing with phytosociology and dynamics of the Tapajós National Forest. Changes in floristic composition over an eight-year period in the logged area, as well as in the undisturbed forest, are discussed.

Material and Methods

The study area

The study area lies in the Tapajós National Forest near kilometer 114 of the Santarém-Cuiabá Road (BR 163) in the municipality of Belterra, State of Pará, Brazil. The Tapajós National Forest area is about 600,000 ha at an altitude *c.* 175 m above the sea level. Its latitude is 2° 40' - 4° 10'S and longitude is 54° 45' - 55° 30'W. The climate is classified by Köppen as Am, which is tropical with one dry season of 2–3 months and an annual rainfall of over 2000 mm. Mean annual air temperature is *c.* 25°C, range 18.4–32.6°C; a

mean relative humidity of 86% (76–93%); a mean annual rainfall of 2110 mm with high rainfall from March to May, and low rainfall from August to November (Carvalho 1982).

The topography is mostly level or slightly rolling. Soil is allic to moderate yellow latosol with heavy clay texture (60–94% of clay), with inclusion of concretionary yellow latosol, derived from clay stone (Brasil-FUPEF 1986). Following the general pattern for the soils of *terra firme* Amazonian forests it is low in nutrients. Dubois (1976) classified the forest type as zonal primary high *terra firme* forest without *babaçu* palm (*Orbignya barbosiana* Burret).

This area was selected because it represents the majority of the *terra firme* dense forests in the region. It is also sufficiently remote from any private settlement to be disturbed by the cutting of trees for building materials or fuel. Apparently the only disturbances in the area were due to hunting of animals for food, collecting edible fruits, tapping of some *Hevea brasiliensis* (HBK) Muell. Arg. trees, and barking trees of medicinal species such as *Aniba canelilla* (Kunth) Mez, *Tabebuia serratifolia* (Vahl) Nicholson and *Stryphnodendron adstringens* Mart. Those activities probably did not cause any serious damage or alter the structure and composition of the forest significantly.

Methods

The research was initiated in 1981 by obtaining a 100% pre-felling forest inventory of trees ≥ 45 cm dbh over a total area of 144 ha. That year

climber cutting was undertaken and the permanent sample plots were established and measured for the first time within the 144 ha. In 1982 the area was logged. The subsequent measurements took place in 1983 - one year after logging, 1987 - five years after logging and in 1989 - seven years after logging. A 36 ha control with permanent plots was established in 1983. It was first measured then and later in 1987 and 1989. The total study area was thus 180 ha.

The experiment was initially designed with four treatments (T1, T2, T3, T4) combining intensities of logging and silvicultural treatments, and a control area (T0). Silvicultural treatments T3 and T4 were not, in fact, applied, because the Brazilian Government did not permit the use of any chemical products in natural forests in the 1980s. Therefore, in the present study, only T1, T2 and T0, totaling 108 ha (36 ha for each treatment and control) are analyzed and discussed.

The forest structure was similar in both the T1 and T2 areas before logging and in the control block (Carvalho, 1992). The Embrapa research team carried out logging. Powersaws were used for cutting. An attempt was made at directional felling and a wheeled skidder was used, operating in trails previously opened by a bulldozer. Particulars of each treatment are as follows:

T1 - this treatment consisted of cutting trees of 38 commercial species ≥ 45 cm dbh. This dbh limit was applied because it is the minimum permitted by the Brazilian Forest Law for felling trees in natural forests. The 38

commercial species were selected from a list of traded species in the regional timber market and were based on their abundance, basal area and timber volume recorded in a previous forest inventory carried out in the study area (Silva *et al.*, 1985).

T2 - cutting of trees of the same 38 commercial species as in T1 but ≥ 55 cm dbh. This diameter limit was applied because generally the sawmills in the Tapajós region are set to use logs ≥ 45 cm diameter. The top diameter of a bole might only measure 45 cm if its dbh is at least 50 cm, although this varies from one species to another. The decision to take 55 cm as the cutting limit assumed that all trees felled would have a diameter greater than 45 cm at any part of their bole. This does not allow for any loss of timber, which occurs, with trees of 45 cm dbh, which is the traditional cutting limit in Amazonia.

T0 - the control area remained in its natural untreated condition.

The statistical design of the treatments constituted four randomized blocks. Each treatment (T1 to T4) was thus replicated four times, with one 9 ha treatment plot per block. The control was a separate block also divided into four plots, each of 9 ha. Each plot was sub-divided into nine 1 ha quadrats. Three of which were selected at random. A 50 m x 50 m permanent sample-plot was set in the center of each. Following the procedure of Silva & Lopes (1984), each permanent sample-plot was further divided into 25 subsample-plots of 10 m x 10 m and marked with permanent top-red-painted stakes. All individuals ≥ 5 cm

dbh were number-tagged, identified and measured. In total, each treatment was applied over 36 ha, and included 3 ha of measured sample-plots.

The guidelines proposed by Synnott (1979) and Whitmore (1989b) for the measurement of the permanent sample-plots in tropical rain forests were followed. Further procedures in measurement were based up on Hutchinson (1982) and Silva & Lopes (1984), described also in Silva (1989)

Individual trees were identified by their vernacular names in the forest. Botanical specimens were collected from less common trees for later identification in the Embrapa Herbarium. Whenever possible, identification was made to the species level (Carvalho, 1992). Species were classified by their wood quality group and ecological class. The wood quality groups were: commercial - species presently sold in the national or international market; potential - species likely to be sold in the near future, mainly on account of their wood properties; non-commercial - species not marketable and without potential, or species whose wood is not yet sufficiently known. This classification was based on the characteristics and uses of wood, the timber market in Brazil and export, according to Brasil-IBDF (1980; 1981; 1988), Brasil-SUDAM (1981), Silva (1989) and Teixeira *et al.* (1988).

The species were further classified as light-demanding or shade-tolerant species, based on the requirement of their seedlings for solar radiation observed casually in the field during the study period (Carvalho, 1992), and according to distinguishing character

suggested by Swaine & Whitmore (1988) and Whitmore (1989a; 1990).

The total number of species, genera and families were determined for T1 and T2 on four occasions and for the unlogged (control) area on three occasions. The numbers of species were also calculated on a per plot basis in order to permit statistical comparison. The family importance values (FIV) were obtained, according to Mori *et al.* (1983), by the sum of family relative values from abundance, dominance and number of species per family ($FIV = FRA + FRD + FRN$, where FRA is the family relative abundance, FRD is the family relative dominance, and FRN is the relative number of species per family). FRA is the percentage of the number of trees ha^{-1} of a certain family in relation to the total number of individuals ($FRA = [\text{number of trees of a family } ha^{-1} / \text{total number of trees } ha^{-1}] \times 100$). FRD is the percentage of basal area of a certain family in relation to the total basal area of the population ($FRD = [\text{basal area of a family} / \text{total basal area}] \times 100$). FRN is the percentage of the number of species of a certain family in relation to the total number of species ($FRN = [\text{number of species of a family} / \text{total number of species}] \times 100$).

Analyses of variance were used to compare the treatments T1 and T2, and logged and unlogged forests. Initially, two-way analyses of variance were performed, as a randomized block design. Since there were no significant differences between the blocks ($p > 0.05$), one-way analyses of variance were applied in each case.

Results

In the 36 permanent sample-plots (totally 9 ha area) 10,926 individuals ≥ 5 cm dbh were recorded. These individuals were the trees standing before logging plus ingrowths to > 5 cm class during the period of studies. Of the total number 83% were identified to the species level, 16.5% to the genus only and 0.5% remained undetermined. There were 257 species in 151 genera from 49 families identified during the study period. Table 1 shows the number of families, genera and species found in the study area. Table 2 shows the changes in number of species over the study period.

Since this study is only concerned with trees ≥ 5 cm dbh, the word "appearance of species" in the results and discussion means the ingrowth of trees into the lowest diameter class. Although individuals of some pioneer species would be able to grow from seed to a tree of 5.0 cm dbh, from one measurement to another, most of the appearance and increases in numbers of species are due to growth of small

trees that already existed. Also disappearance or loss of species means the death of trees ≥ 5.0 cm dbh only.

Analysis of variance, taking into account the mean number of species per plot, showed no significant difference between T1 and T2 during the eight-year study period ($p < 0.05$), but both treatments were significantly different from the unlogged forest ($p < 0.05$), except at the last time of measurement.

In T1 (logging of trees ≥ 45 cm dbh) a new family (Vochysiaceae) appeared one year after logging (1983) but the number of species decreased. The species which disappeared a year after logging were *Ambelania acida* Aubl., *Bagassa guianensis* Aubl., *Bertholletia excelsa* HBK, *Brosimum discolor* Schott, *Cecropia leucoma* Miquel, *Dipteryx odorata* (Aubl.) Willd., *Eperua stipulata* Klein., *Ocotea canaliculata* Mez, *Parkia pendula* Benth., *Phyllanthus nobilis* Muell. Arg., *Saccoglottis guianensis* Benth. and *Vatairea guianensis* Aubl. On the other hand four new species appeared: *Clarisia racemosa* Ruiz. et

Table 1. Numbers of families (NF), genera (NG) and species (NS) of trees ≥ 5 cm dbh in 9 ha sampled at the Tapajós National Forest in Brazilian Amazonia before (1981) and after logging. T1 = logging of trees ≥ 45 cm dbh, T2 = logging of trees ≥ 55 cm dbh.

Year	Logged Forest						Unlogged Forest		
	T1			T2			T0		
	NF	NG	NS	NF	NG	NS	NF	NG	NS
1981	45	128	211	45	133	217	-	-	-
1983	46	126	203	46	129	213	48	137	226
1987	48	134	215	47	133	222	47	135	226
1989	48	137	221	47	135	225	46	133	219

Table 2. Numbers of species of trees ≥ 5 cm dbh over an 8-year period (1981-1989) in 9 ha sampled at the Tapajós National Forest in Brazilian Amazonia. T1 = logging of trees ≥ 45 cm dbh, T2 = logging of trees ≥ 55 cm dbh, T0 = unlogged area. T1 and T2 were logged in 1982. The number of species are per treatment sample (3 ha); the number of species in T1+T2 refers to the sum of sample plots (6 ha) in the logged forest; the number of species in T1+T2+T0 is the total number recorded in the whole sample area (9 ha). T0 was not recorded in 1981.

Treatment	1981	1983	1987	1989
	Numbers of species			
T0	-	226	226	219
T1	211	203	215	221
T2	217	213	222	225
T1+T2	236	228	234	240
T1+T2+T0	236	251	254	257

Pav., *Jacaranda copaia* (Aubl.) D.Don, *Sextonia rubra* (Mez) van der Werff and *Vochysia surinamensis* Stapf. Five years after logging (1987) two new families (Monimiaceae and Chrysobalanaceae) appeared and the numbers of species increased. The new species were: *Aegiphila* sp., *Ambelania grandiflora* Huber, *Cecropia sciadophylla* Mart., *Dimorphandra gardneriana* Jul., *Ficus insipida* Willd., *Parinari barbatum* Ducke, *Pithecellobium* sp., *Protium puncticulatum* Macbride, *Siparuna diciptens* A.DC. and *Vismia japurensis* Reichardt. The species *Eperua stipulata* and *Phyllanthus nobilis*, that were not found immediately after logging (1983), reappeared five years later (1987). No species was lost on this occasion. New changes occurred seven years after logging (1989). *Bagassa guianensis* and *Brosimum discolor*, which disappeared immediately after logging, reappeared this time. The new species seven years after logging were:

Dinizia excelsa Ducke, *Zygia cauliflora* (Willd.) Killip ex Record, *Sapium marmieri* and *Xylopia nitida* Dunal. This time again no species were lost. The changes of species in T1 are shown in Table 3.

In T2 (logging of trees ≥ 55 cm dbh) the families Caricaceae and Connaraceae appeared a year after logging but the Araliaceae disappeared. Seven species disappeared after logging (1983): *Apuleia molaris* Spruce ex Benth., *Bellucia grossularioides* (L.) Triam, *Cariniana micrantha* Ducke, *Schefflera morototoni* (Aubl.) Maguire, *Dinizia excelsa*, *Goupia glabra* Ducke and *Sapium marmieri* Hub. New species this time were *Connarus perrottettii* (D.C.) Planch., *Homalium guianense* (Aubl.) O.Ken. and *Jacaratia spinosa* (Aubl.) DC. Five years after logging (1987) the family Araliaceae reappeared and also the number of species increased. The following new species appeared: *Aegiphila* sp., *Ambelania grandi-*

Table 3. Numbers of species of tree ≥ 5 cm dbh that appeared or disappeared over a 7-year period (1983-1989) after logging at the Tapajós National Forest in Brazilian Amazonia. T1 = logging of trees ≥ 45 cm dbh, T2 = logging of trees ≥ 55 cm dbh, T0 = unlogged forest. T1 and T2 were logged in 1982.

Table 3a. Species that appeared (new species)

Treatment	1983	1987	1989	1983-1989
T0	0	3	3	2
T1	4	14	18	18
T2	3	11	14	14
T1+T2	7	21	26	26

Table 3b. Species that disappeared (lost species)

Treatment	1983	1987	1989	1983-1989
T0	0	3	6	9
T1	12	12	12	8
T2	7	9	9	6
T1+T2	20	22	22	14

Table 3c. Species that appeared and disappeared

Treatment	1983	1987	1989	1983-1989
T0	-	0	1	1
T1	-	2	4	4
T2	-	3	3	3
T1+T2	-	5	8	8

flora, *Capirona huberiana* Ducke, *Eugenia patrisii* Vahl, *Protium puncticulatum*, *Stryphnodendron pulcherrimum* (Willd.) Hochr. and *Vismia japurensis* Reichardt. Species *Bellucia grossularioides*, *Schefflera morototoni* and *Sapium marmieri*, which died a year after logging, reappeared this time. Only *Phyllanthus nobilis* and *Tabebuia impetiginosa* Mart. disappeared at this measurement. Seven years after logging (1989) three new species appeared: *Joannesia*

heveoides Ducke, *Licaria armeniaca* (Nees) Kosterm. and *Protium heptaphyllum* March. There was no loss of species on this occasion. (Tab. 3).

In the unlogged forest, the number of families was reduced between the first (1983) and the second measurements (1987), but the number of species remained the same. The species *Enterolobium sp.*, *Protium sagotianum* March. and *Sclerolobium paraense* Hub. appeared on this occasion, but *Apuleia molaris*, *Parinari barbatum*

and *Swartzia corrugata* Benth. disappeared. At the end of the study period, the species *Annona montana* Macfad., *Erytheca globosa* (Aubl.) Robins., *Schefflera morototoni*, *Diplotropis purpurea* (Rich.) Armsh. and *Xylopia nitida* disappeared. And *Sclerolobium paraense*, which appeared at the second measurement, disappeared this time (Tab. 3).

The 20 most important families in terms of floristic composition, according to the Family Importance Value, are shown in Table 4. These families contain the majority of species and individuals, and those of higher basal areas. They are also responsible for almost all the changes in floristic composition over the eight-year period. Most of the species that disappeared, as well as those that reappeared and the new species belong to this families.

Comparing the two intensities of logging in Tapajós, the number of species that disappeared over the study period in T1 (8 species) was slightly higher than the corresponding number in the less intense T2 (6 species). Eighteen new species appeared in T1 and fourteen in T2. The numbers of species that disappeared and then reappeared over the period, four in T1 and three in T2, were almost identical. In general the changes in T1, where there was more solar radiation because of the greater intensity of logging, were a little higher than in T2. In addition to the slight differences in changes of species composition that occurred between the two intensities of logging, between logged and unlogged forest, differences were also

found within treatments. For example, reduction in numbers of species caused by logging ranged from 3 to 19 species per plot in T1 and from 0 to 14 species per plot in T2. Corresponding coefficients of variation were 46% and 48%. In general, the greatest reduction in the number of species occurred in plots where more trees were cut or destroyed by logging, and where high basal areas were reduced.

One plot in T1, which by chance had no tree cut, showed a reduction of three species due to 36 small individuals destroyed by logging operations. But in T2, a plot apparently not physically disturbed during exploitation had no lost species. Moreover that same plot in T1 showed an increase of 9 species during the eight-year period, while the one in T2 increased by only two species.

The changes were smaller in the unlogged forest. Although the number of species (nine) that disappeared over the study period was higher than in T1, the number of new species (only two) was 11% lower than in T1 and 14% lower than in T2. One species (*Sclerolobium paraense*) appeared and disappeared over the period. Analyzing each plot separately, slight differences are evident. For example, some plots gained new species over the study period without any loss and vice versa. No plot kept the same number of species over the entire period. In the unlogged area the range in net number of species was 0-7. It was observed that natural disturbances such as fall or death of neighboring trees of the measurement plot, generated an increase in amount of solar radiation

Table 4. Family Importance Value (FIV) of the 20 most important families in 9 ha sampled at the Tapajós National Forest, selected according to numbers of genera and species of trees ≥ 5 cm dbh, before logging (1981) and seven years after logging (1989) in the logged area, and for the first (1983) and third (1989) measurements in the unlogged forest. T1 = logging of trees ≥ 45 cm dbh, T2 = logging of trees ≥ 55 cm dbh.

Family	T1		T2		Unlogged Forest	
	1981	1989	1981	1989	1981	1989
Leguminosae	42.7	51.1	50.7	51.3	53.1	49.9
Sapotaceae	24.9	22.1	27.1	23.1	29.4	30.4
Lecythidaceae	26.5	21.1	26.5	19.1	22.1	23.1
Moraceae	16.8	28.6	18.6	24.6	15.6	16.6
Violaceae	21.2	21.2	20.2	19.2	15.6	15.6
Burseraceae	11.9	11.3	14.9	15.1	14.3	14.7
Annonaceae	12.3	10.7	9.7	9.7	13.7	12.9
Lauraceae	13.4	12.4	14.8	16.2	12.4	12.4
Apocynaceae	7.7	5.7	11.1	10.5	12.1	12.1
Meliaceae	14.9	8.9	11.6	8.6	6.6	6.6
Myristicaceae	6.6	6.6	6.6	5.6	6.6	6.6
Rubiaceae	6.3	5.3	4.9	5.3	6.3	6.3
Myrtaceae	3.9	3.9	3.6	3.9	5.9	5.9
Euphorbiaceae	6.3	5.7	4.7	4.7	4.9	4.9
Guttiferae	3.9	4.3	2.5	2.9	4.3	4.3
Sterculiaceae	3.2	3.2	3.2	3.2	4.2	4.2
Anacardiaceae	2.5	2.5	3.2	2.2	3.6	3.6
Melastomataceae	3.3	3.3	3.3	3.3	2.9	2.9
Flacourtiaceae	3.6	4.6	2.5	3.9	2.5	2.5
Olacaceae	4.4	5.4	6.4	7.4	2.4	2.4

inside the plot, allowing the appearance of some species and an increase in abundance of others.

Comparing the species themselves, those that disappeared in T1 were different from those in T2. Only *Phyllanthus nobilis* coincided. The species that appeared immediately after logging were different between

treatments. Five years after logging four (17%) new species were common to both treatments T1 and T2. But again in the following years all species that changed were different between the treatments. In general, most of the species involved in these dynamics in both treatments belong to the group of non-commercial species, according to

the classification given in Carvalho (1992).

The loss of species in the logged area in Tapajós occurred mainly because of the logging. Of the loss in T1 and T2, 38% and 44% respectively, were due to commercial species that were cut. Non-commercial species were lost in consequence of the logging operations. Light-demanding species, including the three wood quality groups, in both treatments, represented 55% of the lost species.

The composition of new species in the logged area was different between the two treatments in some respect, when comparing groups of species. In T1, 51% of the 18 new species were light-demanding against 47% of 14 new species in T2. But in T1 23% were commercial species and 14% were potential and 64% were non-commercial species against T2 where 6% were commercial and 94% were non-commercial. Apparently the most intense logging (T1) was more favorable to growing young trees of commercial (mainly light-demanding) species during the eight-year period.

In addition, comparing the gain and loss of species between T1 and the unlogged forest, no species coincided. The only co-incidences noted were when the loss of species was compared between T2 and the unlogged forest. Two species (*Apuleia molaris* and *Schefflera morototoni*) disappeared during the study period in both areas.

In the unlogged forest, the loss of species was due to natural causes. *Schefflera morototoni* and *Diptlotropis purpurea* disappeared

when the only two big trees (>60 cm dbh) fell in storms. The other disappeared through death of individuals <15 cm dbh, probably in consequence of the fall of those large trees. Fifty-six per cent of the lost species were shade-tolerant. Of the three new species that came in, two were shade-tolerant. Considering wood quality, two species were non-commercial (*Protium sagotianum*, *Sclerolobium paraense*) and one was potentially commercial (*Enterolobium sp.*).

Discussion

These changes can be compared with those observed in other tropical forests. Although there are similarities at family level between the tropical rain forests worldwide, the genera and species are almost entirely different, even within each rain forest region (Whitmore, 1990). Differences in floristic composition are also noted between places in the same area at the same time. This was evident in the treatments of the present study and was shown as well in three treatment plots analyzed by Okali & Ola-Adams (1987) in Omo Forest Reserve, Nigeria. Further evidence of these differences can be found by comparing the studies done by Black *et al.* (1950), Pires *et al.* (1953), Costa *et al.* (1999), Soares & Carvalho (1999), Lima & Carvalho (2000) and Sandel & Carvalho (2000) on six areas in the State of Pará in Brazilian Amazonia. Differences are more evident between distant areas, for example between countries or continents. In a study review on dynamics of 18 tropical sites,

including America, Africa and South-east Asia, considering trees ≥ 10 cm dbh, Swaine *et al.* (1987b) found a wide variation in number of species, from 52 to 141 ha⁻¹.

In 2 ha of moist semi-deciduous forest at Kade, Ghana, Swaine *et al.* (1987a) identified 120 species from 36 families, over a 14-year period. Seven species disappeared but were replaced by seven new species, representing a turnover of about 1% annually. They enumerated trees and climbers ≥ 9.5 cm dbh. In the present study there were 201 species, considering only trees ≥ 9.5 cm dbh in a 2 ha sample. This number is higher than that found at Kade, in a time period longer than that in Tapajós.

Probably because eight years is not a long period, the floristic composition was not highly dynamic. In the logged forest, 14 species disappeared over the 8-year period, but 26 new ones appeared. In the unlogged area, nine species disappeared and only two new ones came in. In the study of Manokaran & Kochummen (1987) in an unlogged forest in Peninsular Malaysia in a 34-year period, 48 species disappeared and were replaced by another 48; and six species appeared after the beginning of the period and disappeared before the end. In Tapajós seven species disappeared and reappeared during the period. In this Malaysian study the pioneer species, defined according to the classification given by Swaine & Whitmore (1988) and discussed in Whitmore (1989a, 1990), were not very abundant and declined in numbers of species and

trees from the beginning to the end of the period. Also in Tapajós, in the unlogged forest, the abundance of pioneer species was very low and remained almost the same during the whole period, varying only by 0.3% from the beginning to the end. On the other hand, in the logged area, 17 new pioneer species appeared over the study period. The abundance of the majority of the pioneer species increased each year after logging, but this is not a worrying problem because most of them are often short-lived, according to characteristics described by Swaine & Whitmore (1988), and others are commercial species such as *Caryocar villosum* (Aubl.) Pers., *Cordia alliodora* (Ruiz & Pav.) Oken, *Cordia bicolor* A.DC and *Jacaranda copaia*, or potential species such as *Bixa arborea* Huber, *Laetia procera* (Poepp. et Endl.) Eichl., *Tachigali chrysophylla* (Poepp.) Zarucchy & Herend. and *Tachigali myrmecophila* (Ducke) Ducke, amongst others. Abundance of species is discussed in detail in Carvalho (1992).

Some ecological characteristics and effects of logging in a rain forest in north Queensland were conducted by Nicholson *et al.* (1988). They analyzed tree species ≥ 10 cm dbh in 13 plots of different sizes, under different silvicultural interventions and in different sites. Plot 408 was the largest (1.6 ha) and the richest with 125 species before logging. This number is the same as that found in only 0.25 ha of unlogged forest in Tapajós also considering trees ≥ 10 cm dbh. Co-incidentally the number of species after log-

ging in plot 408 in Queensland (117 species) was identical to that recorded one year after the heavier logging in only 0.25 ha in Tapajós. The losses of species in most of the study plots in Queensland were higher than in Tapajós, probably because the logging was also more intense and because the logging reported was not the first cut (Saxon, 1990).

In studies that deal with various plant life forms the numbers of species are markedly higher. For instance, in a 12.4 ha area at La Selva, Costa Rica, Lieberman & Lieberman (1987) found 269 species in 162 genera and 62 families, considering individuals ≥ 10 cm dbh. The species measured included: seven palms, twenty lianas, two strangler figs, one tree-fern and 239 true trees. Considering only true trees, the number of species recorded in that study was the same as found in only 6 ha in Tapajós, taking into account trees ≥ 10 cm dbh.

The present study also indicates once more that the intensity of logging has an important influence on floristic composition, although it is not significant in terms of numbers of species per area. This was noted by Whitmore (1984) in relation to the size of gaps in natural forests: different species are successful in growing up in gaps of different size. Whitmore (1988) further remarked that the species composition in exploited areas depends on what grows up in the gaps created by logging, and that the greater the disturbance the more the light-demanding species are favored. This was noted in two heavily logged areas studied by Enright (1978)

in Papua New Guinea, and by Johns (1988) at Sungai Tekam, West Malaysia, where the colonizing tree species, mainly Euphorbiaceae, germinated and grew rapidly immediately after logging, invading all big gaps.

Nicholson *et al.* (1988) in Queensland, found that logging reduced species numbers among trees ≥ 10 cm dbh, but they also showed that there was an increase in numbers of species with time after logging. The present study in Tapajós revealed that the number of species was reduced immediately after logging but it increased with time (Tab. 2) and showed more species seven years after logging (1989 - end of the study period) than before (1981). It is not known yet how long the number of species will increase after logging, or when it will start to decline. However it is especially worthy of attention that in the logged forest in Tapajós, considering trees ≥ 5 cm dbh or trees ≥ 10 cm dbh, the number of species rose with time and that there were also commercial species among these new ones.

Nicholson *et al.* (1988) stated that in some virgin rain forests in Queensland there is a decrease in number of species with time, but the re-examination of their data by Saxon (1990) showed that this occurred only in two plots, the largest and the smallest, while in another two, which suffered natural disturbance, it did not happen. Saxon concluded that in presence of natural disturbances species richness and species diversity did not decline in virgin forest. Nicholson *et al.* (1990) replied that if a natural disturbance is similar to a moderate or

heavy logging no distinction can be made between natural and artificial disturbance about their response to changes in numbers of species and diversity. In Tapajós there was a loss of 3% in numbers of species in the unlogged forest over seven years as already discussed (Tabs 2, 3). But due to the short study period this fact gives no evidence that the number of lost species will continue increasing with time. Any loss of large trees and consequent increase of solar radiation could change that tendency in the near future.

To conclude, changes in floristic composition occurred under undisturbed conditions as well as in response to natural or man-made disturbances. The number of species decreased immediately after logging, but started to increase by the fifth year after logging. The more heavily logged plots responded more to disturbance, as judged by the increase in the number of species during the period following logging. It appears that the forest can rapidly recover its initial floristic composition without intervention, agreeing with statements made by Carvalho (1997) on dynamics of natural forests and its implication to the forest management.

Literature cited

- Black, G.A.; Dobzhansky, T.H.; Pavan, C. 1950. Some attempts to estimate species diversity and population density of trees in Amazonia forests. *Botanical Gazette*, v.111, p.413-425.
- Carvalho, J.O.P. de. 1982. *Análise estrutural da regeneração natural em floresta tropical densa na região do Tapajós no Estado do Pará*. Curitiba: UFPR. 129p. Dissertação Mestrado.
- Carvalho, J.O.P. de. 1984. *Manejo de regeneração natural de espécies florestais*. Belém: EMBRAPA-CPATU. 22p. (EMBRAPA-CPATU. Documentos, 34).
- Carvalho, J.O.P. de. 1987. *Subsídios para o manejo de florestas naturais na Amazônia brasileira: resultados da pesquisa da Embrapa/IBDF-PNPF*. Belém: EMBRAPA - CPATU. 35p. (EMBRAPA - CPATU. Documentos, 43).
- Carvalho, J.O.P. de. 1992. *Structure and dynamics of a logged over Brazilian Amazonian rain forest*. Oxford: University of Oxford. 215p. D.Phil. Thesis.
- Carvalho, J.O.P. de. 1997. *Dinâmica de florestas naturais e sua implicação para o manejo florestal*. Curitiba: EMBRAPA-CNPQ. p43-55. (EMBRAPA-CNPQ. Documentos, 34).
- Carvalho, J.O.P. de; Silva, J.N.M.; Lopes, J. do C.A. 1993. Spatial distribution of tree species of a terra firme rain forest in Brazilian Amazon. In: Congresso Florestal Panamericano, 1.; Congresso Florestal Brasileiro, 7. 1993, Curitiba, PR. *Floresta para o desenvolvimento: política, ambiente, tecnologia e mercado – anais*. Curitiba: SBS/SBEF. v.1, p.393-396.
- Carvalho, J.O.P. de; Silva, J.N.M.; Lopes, J. do C.A.; Jimenez Valcarcel, V.M.; Graaf, N.R. de. 1986. Redução da densidade de uma floresta tropical úmida densa devido à exploração mecanizada. In: Symposium of the Humid Tropics, 1., 1984, Belém, PA. *Proceedings*. Belém: EMBRAPA-CPATU. v.2, p.269-281. (EMBRAPA-CPATU. Documentos, 36).
- Carvalho, J.O.P. de; Silva, J.N.M.; Lopes, J. do C.A.; Costa, H.B. da. 1984. *Manejo de florestas naturais do trópico úmido com referência especial à Floresta Nacional do Tapajós no Estado do Pará*. Belém: EMBRAPA-CPATU. 14p. (EMBRAPA-CPATU. Documentos, 26).
- Costa, D.H.M.; Ferreira, C.A.P.; Silva, J.N.M.; Lopes, J. do C.A.; Carvalho, J.O.P. de. 1999. *Potencial madeireiro de floresta densa no município de Moju no Pará*. Belém: EMBRAPA-CPATU. p.138-141.

- (EMBRAPA-CPATU. Documentos, 123).
- Costa, H.B. da; Lopes, J. do C.A. 1982. Rendimento de derruba e extração mecanizada em florestas de terra firme da Amazônia. In: Congresso Nacional sobre Essências Nativas, 2., 1982, Campos do Jordão. *Anais*. São Paulo: Instituto Florestal; *Silvicultura em São Paulo*, v. 16 A, p.1253-1258, 1982.
- Dubois, J.C.L. 1976. *Preliminary management forest guidelines for the National Forest of Tapajós*. Belém: IBDF/PRODEPEF. 42p.
- Enright, N. J. 1978. The effects of logging on the regeneration and nutrient budget of a *Araucaria cunninghamii* dominated tropical rainforest in Papua New Guinea. *The Malaysian Forester*, v.41, p.303-318.
- Fundação de Pesquisas Florestais do Paraná (Curitiba, PR). 1986. *Relações entre solos e a vegetação natural em área da Floresta Nacional do Tapajós*. Curitiba.
- Hutchinson, I.D. 1982. *Field enumeration of permanent sample plots in the mixed Dipterocarp forest of Sarawak*. [S.l.]: FAO. (FAO. Report FO:MAL/76/008. Field Documents, 16).
- Instituto Brasileiro de Desenvolvimento Florestal (IBDF). 1988. *Amazonian timbers: characteristics and utilization*; Curuá-Una Experimental Forest Station. Brasília, 236p.
- Instituto Brasileiro de Desenvolvimento Florestal (IBDF).. 1981. *Amazonian timbers: characteristics and utilization*; Tapajós National Forest. Brasília. 113p.
- Instituto Brasileiro de Desenvolvimento Florestal (IBDF).. 1980. *Woods from Tucuruí: characteristics and application*. Brasília. 49p.
- Johns, A.D. 1988. Effects of selective timber extraction on rain forest structure and composition and some consequences for frugivores and folivores. *Biotropica*, v.20, p.31-37.
- Lieberman, D.; Lieberman, M. 1987. Forest tree growth and dynamics at La Selva, Costa Rica (1969-1982). *Journal of Tropical Ecology*, v.3, p.347-358.
- Lima, S.F. de; Carvalho, J.O.P. de. 2000. *Estrutura de uma floresta de terra firme na região de Marabá-PA: a posição do mogno em relação às outras espécies da comunidade*. Belém: EMBRAPA-CPATU. 5p. (EMBRAPA-CPATU. Comunicado Técnico, 22).
- Lopes, J. do C.A. 1993. *Demografia e flutuações temporais da regeneração natural após uma exploração florestal*: Flona Tapajós-PA. Piracicaba: USP. 133p. Dissertação Mestrado.
- Lopes, J. do C.A.; Carvalho, J.O.P. de; Silva, J.N.M.; Costa, H.B. da. 1986. Sociabilidade entre 18 espécies comerciais ocorrentes na Floresta Nacional do Tapajós. In: Symposium of the Humid Tropics, 1., 1984, Belém, PA. *Proceedings*. Belém: EMBRAPA-CPATU. v.2, p. 263-268. (EMBRAPA-CPATU. Documentos, 36).
- Manokaran, N.; Kochummen, K.M. 1987. Recruitment, growth and mortality of tree species in a lowland dipterocarp forest in Peninsular Malaysia. *Journal of Tropical Ecology*, v.3, p.315-330.
- Mori, S.A.; Boom, B.M.; Carvalho, A.M. de; Santos, T.S. dos. 1983. Southern Bahian moist forests. *The Botanical Review*, v.49, p.155-232.
- Nicholson, D.I.; Henry, N.B.; Rudder, J. 1990. Reply-disturbance regimes in North Queensland rainforests: a re-evaluation of their relationship to species richness and diversity. *Australian Journal of Ecology*, v.15, p.245-246.
- Nicholson, D.I.; Henry, N.B.; Rudder, J. 1998. Stand changes in north Queensland rainforests. *Proceedings of the Ecological Society of Australia*, v.15, p.61-80.
- Okali, D.U.U.; Ola-Adams, B.A. 1987. Tree populations changes in treated rain forest at Omo Forest Reserve, South-western Nigeria. *Journal of Tropical Ecology*, v.3, p.291-313.
- Pires, J.M.; Dobzhansky, T.H.; Black, G.A. 1953. An estimate of the number of species of trees in an Amazonian forest community. *Botanical Gazette*, v.114, p.467-

477.

- Sandel, M.P.; Carvalho, J.O.P. de. 2000. *Composição florística e estrutura de uma área de cinco hectares de mata alta sem babaçu na Floresta Nacional do Tapajós*. Belém: Embrapa Amazônia Oriental, 15p. (Embrapa Amazônia Oriental, Documentos, 63).
- Saxon, E.C. 1990. Disturbance regimes in North Queensland rainforests: a re-evaluation of their relationship to species richness and diversity. *Australian Journal of Ecology*, v.15, p.,241-244.
- Silva, J.N.M. 1989. *The behaviour of the tropical rain forest of the Brazilian Amazon after logging*. Oxford: University of Oxford. D. Phil. Thesis.
- Silva, J.N.M.; Lopes, J. do C.A. 1984. *Inventário florestal contínuo em florestas tropicais: a metodologia utilizada pela EMBRAPA-CPATU na Amazônia brasileira*. Belém: EMBRAPA-CPATU. 36p. (EMBRAPA-CPATU. Documentos, 33).
- Silva, J.N.M.; Carvalho, J.O.P. de; Lopes, J. do C.A. 1985. Inventário florestal de uma área experimental na Floresta Nacional do Tapajós. *Boletim de Pesquisa Florestal*, v.10/11, p.38-110.
- Soares, M.H.M.; Carvalho, J.O.P. de. 1999. *Diversidade florística em uma área de 200 hectares de floresta natural no município de Moju no Pará*. Belém: EMBRAPA-CPATU. p.110-115. (EMBRAPA-CPATU. Documentos, 123).
- SUDAM (Belém, PA). 1981. *Grupamento de espécies tropicais da Amazônia por similaridade de características básicas e por utilização*. Belém.
- Swaine, M.D.; Hall, J.B.; Alexander, I.J. 1987a. Tree population dynamics at Kade, Ghana (1968-1982). *Journal of Tropical Ecology*, v.3, p.331-345.
- Swaine, M.D.; Lieberman, D.; Putz, F.E. 1987b. The dynamics of tree populations in tropical forest: a review. *Journal of Tropical Ecology*, v.3, p.359-366.
- Swaine, M.D.; Whitmore, T.C. 1988. On the definition of ecological species groups in tropical rain forests. *Vegetatio*, v.75, p.81-86.
- Synnott, T.J. 1979. A manual of permanent plots procedures for tropical rain forests. *Tropical Forestry Papers*, v.14.
- Teixeira, D.E.; Santana, M.A.E.; Souza, M.R. de. 1988. *Amazonian timbers for the international market*. Brasília: IBDF/ITTO. (ITTO. Technical Series, 1).
- Whitmore, T.C. 1984. *Tropical rain forests of the Far East*, 2.ed. Oxford: Clarendon. 352p.
- Whitmore, T.C. 1988. The influence of tree population dynamics on forest species composition. In: Davy, A.; Hutchings, M.J. (eds.) *Plant population ecology*. Oxford: Blackwell. p.271-291.
- Whitmore, T.C. 1989a. Canopy gaps and the two major groups of forest trees. *Ecology*, v.70 p.536-538.
- Whitmore, T.C. 1989b. Changes over twenty-one years in the Kolombangara rain forests. *Journal of Ecology*, v.77, p.469-483.
- Whitmore, T.C. 1990. *An introduction to tropical rain forests*. Oxford: Clarendon. 226p.

Aceito para publicação em 22/04/2002